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DESTRUCTION OF VOLATILE ORGANIC COMPOUNDS BY CATALYTIC OXIDATION

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Abstract

Volatile organic compounds, emitted by a lot of sources, particularly from gasoline engines of automobiles, are air pollutants that give rise to deleterious health and environmental effects. Total catalytic oxidation is an attractive method in controlling these emissions due the great amounts of energy saved, by the moderate temperature involved. For wide implementation of catalytic combustion, thermally, mechanically and chemically stable catalysts are required. The paper is a short reviewing of the available catalysts and also of the new innovations concerning either enhance of their performances or new better formulations.

Keywords: volatile organic compounds, catalytic oxidation, catalyst life

1. What are VOCs and why have to eliminate them

VOCs are everywhere: at homes, at dry cleaning units, at food processors, at petroleum refineries, at airports and automobile service stations, at electronics and chemical plants, at print and paint shops etc. VOCs are those chemicals that exposed to air volatilize readily. So, these compounds contribute to air pollution by their vapour and to water and soil pollution by that fraction which can be carried deep into the soil by rainwater and melting snow. This work studies VOCs only from the viewpoint of the air pollution.

1.1. Environment and VOCs

VOCs are dangerous for humans, fauna and flora. Firstly, these compounds have been proved to cause a variety of adverse health effects (Minnesota Department of Health, 2003, Regional Environmental Assessments, 2002; U.S. EPA, 2003). Generally, allergic skin reaction, dizziness, headaches, eye and respiratory tract irritation, coughing, visual disorders, memory impairment, confusion, anemia and fatigue are some of the symptoms associated with short-term exposure. Kidney, liver, brain damage, cardiac sensitization reactions and also damage to the nervous, reproductive and immune systems are caused by long-term exposure. VOCs are also "endocrine impostors". Some are

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also known or suspected of causing cancer. Also, the natural plant processes can be destroyed or interrupted by some organic volatile compounds. In the atmosphere these chemicals are transformed in other compounds even more dangerous for humans, fauna and flora. The ground level or "bad" ozone, a key ingredient of urban smog, is created by a chemical reaction between oxides of nitrogen (NOx) and volatile organic compounds in the presence of sunlight. Repeated exposure to ozone pollution may cause a variety of health problems (Clean Air Counts, 2003); ozone also reduces the growth rate of trees and agricultural production. VOCs participate in the greenhouse effect, acid rains and stratospheric or "good" ozone depletion, which are serious environmental problems (Avila et al., 1998). Consequently, VOCs are a great group of air pollutants and our duty is to eliminate them from being released. For this reason, firstly, we must know who are they and which are their sources.

1.2. VOCs and their sources

The USA Environmental Protection Agency (EPA) gave the following definition for VOCs – any compound of carbon, excluding carbon monoxide, carbon dioxide, carbonic acid, metallic carbides or carbonates, and ammonium carbonate, which participates in atmospheric photochemical reactions. This includes any organic compound other than the compounds that have been determined to have negligible photochemical reactivity (e.g. methane, ethane, methylene chloride...). According to this definition a wide variety of organic compounds, such as aliphatic, aromatic and chlorinated hydrocarbons, aldehydes, ketones, esters, organic acids and alcohols can be considered as VOC (see the complete list in (Ron Joseph & Associates, 2000)).

Emissions of VOCs could appear from natural sources, but most VOCs emissions result from man-made sources. Approximately 235 millions tonnes of VOCs are released per year into the atmosphere from anthropogenic sources (Guenther et al., 1995) and that is because there are thousands of different VOCs produced and used in our daily lives. In the USA ca. 40% of the VOCs emissions are released from transportation activities and the remaining 60% result from stationary sources. The last sources are about equally divided between fuel combustion, industrial manufacturing and solvent emissions (Horsley, 1993) and might be interior or exterior sources. Some examples of commercial operations that produce volatile organic compounds emissions are: chemical plants, petroleum refineries, pharmaceutical plants, automobile manufacturers, airplane manufacturers, food manufacturers. textiles manufacturers, printing plants, can coating plants, wire enameling plants, electronic component plants, painting facilities, wood stoves. In all these manufacturing processes organic materials are present as petroleum derivatives, chemicals, solvents, pigments, release agents, coatings, paints, adhesives. Usually they leave the plant in a form of a gaseous effluent that contains low concentrations of organics and that is vented into the atmosphere. Concerning household, some products that contain volatile organic compounds and pollutes the indoor air are: paints, paint strippers, lacquers, solvents, wood preservatives,

aerosol sprays (cosmetics), cleaners and disinfectants, moth repellents and pesticides, air fresheners, stored fuels and automotive products, hobby supplies, dry-cleaned clothing, building materials and furnishings, office equipment such as copiers and printers, correction fluids and carbonless copy paper, permanent markers, photographic solutions and so forth (Regional Environmental Assessments, 2002; U.S. EPA, 2003; Minnesota Department of Health, 2003).

1.3. Measures against VOCs emissions

The progressive increase of VOCs emissions, as well as of the information about their harmful effects, determined environmental legislation of the worldwide to introduce more stringent regulations to control their release (Clean Air Counts, 2003; U.S. EPA, 2003;). In United States of America, the champion in this fight against the environment pollution, *The Clean Air Act Amendments of 1990* calls for a 90% reduction in emissions of 189 toxic chemicals, 70% of them VOCs, over the next nine years (Parkinson, 1991). For that call, USA *Environmental Protection Agency (EPA)*, which has been working for a cleaner and healthier environment for the American people for 30 years, states and cities implemented programs to further reduce emissions from stationary and mobile sources. Some solutions are green technologies, cleaner car and fuel, gas station using special nozzles at the pumps to recapture vapour, vehicle inspection programs, etc. At the beginning of 1997 there are over 6000 systems operating throughout the world controlling VOC emissions (Anguil Environmental Systems Inc.).

The methods used for decreasing emissions of VOCs can be divided (Dumitriu and Hulea, 1997) in: optimization of the process by implementation of "green technologies "and cleaning of flue gases. Many different techniques, which can be the "end-of-the-pipe" control technologies for VOCs abatement, are available (GlobalSpec Inc., 2003; Haslego, 2002; Uberoi, 2000) They use either the capture of the pollutants or/and their destruction. When products recovery is desired capture techniques like as *condensation, absorption* and *adsorption* are required. Contaminants can be destroyed by a flame (*incineration*) or by a microorganism (*biodegradation*). The selection of the most appropriate method will depend on following criteria: desired control efficiency, contaminant properties (e.g. stickiness, volatility, molecular weight, pH, solubility in water or other solvents), air stream properties (concentration, flow rate, heat content, vapour pressure, temperature, particulate matter loading, moisture content), safety issues (flammability, explosivity), value of recovered material etc.

Among the technologies mentioned above for VOCs abatement the most widely implemented is the *incineration (oxidation)*. Destruction efficiency of this method can achieve more than 99, 99% depending on the type of the system (GlobalSpec Inc., 2003). There are two forms of incineration: *thermal oxidation* and *catalytic oxidation*.

2. Catalytic oxidation versus thermal oxidation as methods for reduction of VOCs emissions

By total oxidation, VOCs molecules are transformed into CO₂ and H₂O. For break apart VOCs molecules and oxidize them to carbon dioxide and water, that means for overcome the activation energy of this reaction, a certain amount of energy is required, so the temperature of the air stream containing VOCs molecules have to be raised to a certain point - typically in excess of 760°C (Cooley, 2002). This is a thermal oxidation and it occurs in a thermal oxidizer (Auckland Regional Council, 2000; Anguil Environmental Systems Inc.) involving a high temperature system (e.g. direct flames, plasma torch and indirectly heated ceramic bed oxidizer). It is important to know that this reaction is an exothermic one, so the resulted heat could be used to preheat the incoming exhaust. Thus, the input energy required is less and fuel costs are reduced. The heat recovery may be achieved by a recuperative or a regenerative system. In the first the waste heat is transferred to the process air through a surface (air-toair exchanger, for example) as long as a temperature gradient exists between the two fluids. In the second one the waste heat is transferred to the surface of a ceramic stoneware over time, after which the incoming air passing over the same surface absorbs the heat.

Another way to reduce the input energy and thus the costs with fuel is the *catalytic oxidation*. In this case, the same reaction occurs into a catalyst bed and the activation energy barrier is lower, so the temperature required is significantly lower – typically starting at 205 to 260 °C (Cooley, 2002). In addition, the catalyst facilitates a faster reaction in smaller volume equipment.

Various catalytic technologies of incineration have been developed by different companies. For example, *Anguil Environmental Systems Inc.* is a leader concerning the technologies used in controlling air pollution (Anguil Environmental Systems Inc.). They make thermal and catalytic incinerators, also.

Concerning the technological aspects, both thermal oxidation and catalytic oxidation have limitations. In choosing the right technology, a careful review of all the aspects needs. In (Auckland Regional Council, 2000; Uberoi, 2000) are described the "suitable incinerator applications" and "recommended performance and design criteria for incinerators". Factors, which must be careful assessed when VOCs abatement by catalytic oxidation occurs, are (Cooley, 2002): the lower explosive limit – for safety considerations the gas stream (inlet), which is introduced with the combustion air, must have its concentration less than 25% of the lower explosive limit (LEL), the flow rate – typical system capacities range from 2,8 to 283 cubic meters per minute and VOCs concentration – catalytic oxidation typically has environmental applications for VOCs concentrations "from a few hundred to a few thousand parts per million (ppm) but, is the preferred technology for "relatively low VOCs concentrations (~10 ppm v)".

Do not forget the three "T's" of combustion: turbulence, time and temperature.

Catalytic oxidation has been applied for odour control (Hermia, 1993) and treatment of emissions containing evaporated solvents (Vigneron, 1996). Also, a technology for treatment of very low concentration VOCs in indoor air is being sought (O'Malley and Hodnett, 1999).

A great advantage of catalytic oxidation is that it can be also used for mobile sources. It is known that automotive exhaust gases formed in the gasoline engines contain many environmentally harmful compounds, as a result of incomplete combustion, like carbon monoxide (CO), hydrocarbons (HC) and NO_x. The developing of an interesting device called catalytic converter, which treats the exhaust gas before it leaves the car, is a real progress (Nice, 2003). This device converts the three regulated emissions - CO, VOCs and NO_x (it is known like *Three Way Catalyst- TWC*) at CO₂, H₂O and N₂. TWCs have been fitted to cars since the 1970s (Heck and Farrauto, 2001) and due to the increased demand for low-emission vehicles (see Euro Standard) (Vehicle Certification Agency, 2002), most automobiles are now supplied with a three-way catalytic converter (85 % of all new cars world wide are fitted with TWC). There is a short history of the evolution of the converters since 1976 up to now (Dumitriu and Hulea, 1997).

The application of catalytic oxidation like a method for emission control is described in the literature in several review articles (Prasad, 1984; Nakajima, 1991). The moderate temperatures involved into a catalytic oxidizer cause, besides the attractive saved energy, the possibility to use heat exchangers made by common alloys (John Zink Company, 1991) and also a higher selectivity to CO_2 and H_2O ; harmful by-products, such as dioxin and nitrogen oxides, which are formed into thermal oxidizers (Ramachandran, 1996) (involve high temperatures), are not generated. It has been reported that formation of partially oxidized intermediates, much toxic than the originals could occur (Lintz and Wittstock, 2001).

The economic factors, which have made catalytic oxidation an attractive method, are in competition with some negative factors such as the price of the catalyst and its lifetime. The last factor is depicted in the paragraph 3.3 (Catalyst life).

3. About catalytic oxidizers and catalytic systems

3.1. The catalytic oxidizer

A catalytic oxidizer is similar in design to a recuperative one (Auckland Regional Council, 2000; Anguil Environmental Systems Inc.), except that the oxidation of VOCs occurs into the catalyst bed instead into the open flames. The air containing VOCs is ducted to the catalytic system passing through a heat exchanger for preheating. Into the burner chamber the air is heated to the specific VOC catalyst reaction temperature. The heat released due to oxidation

process preheats the incoming exhaust. The difference between the thermal oxidizer and the catalytic one is that the catalyst allows the same VOC destruction at considerably lower temperature, saving fuel costs. The scheme of a catalytic oxidizer made by *Anguil Environmental Systems Inc.* is given in Fig. 1.



Fig. 1. Catalytic oxidizer (Anguil Environmental Systems Inc.).

3.2. The configuration of the catalytic system

A typical catalytic system is composed of a catalytic active material, finely dispersed across an extended surface provided by a mechanical support, which has been previously washcoated with a highly porous material (Cooley, 2002; John Zink Company, 1991). The active material presents the "sites" where the reaction between oxygen and VOCs occurs. Noble metals and certain metal oxides are the most active materials for VOCs oxidation (Lahousse et al., 1998b). The washcoat with its high surface area maximizes the number of active sites available to the reactants and it is called "the support material". The physical structure and the mechanical integrity of the ensemble are provided by the "mechanical support", which takes several forms such as pellet or monolith honeycomb, by metal (stainless steel) or ceramic. Dumitriu and Hulea (1997) have made a depiction of all these systems. Typically, manufacturers employ the ceramic honeycomb, with a critical design in achieving maximum destruction removal efficiency (DRE) (Cooley, 2002). In Fig. 2 it is presented the way of distribution of the washcoat on the monolith walls.

3.3. Catalyst life

Limitations of catalytic oxidation are affined with the limitation of the catalytic system: catalyst life and catalyst cost.

Theoretically, the catalyst is ageless. In practice its life is reduced (Cooley, 2002). It is known that the catalytic activity of a catalyst in VOCs complete oxidation studies is evaluated by tracing the ignition (light-off) curves

("VOC conversion versus reaction temperature") (John Zink Company, 1991; Paulis, 2000). After some period of operation this curve may indicate a loss of activity (deactivation). The three main reasons for catalyst deactivation are *poisoning, fouling and thermal deactivation*, which can act alone or in combination. The followings are for explain away the reasons for these phenomena.



Fig. 2. Monolith ceramic – section (Dumitriu and Hulea, 1997).

The active sites of a catalyst could react with a second chemical and if the catalytic activity is enhanced, the phenomenon is known as *doping*, if decreased, as poisoning (Thomas and Thomas, 1967). There are some chemical species contained in exhaust gases, which are considered like poisons: halogens, silicones, phosphorus, sulphur, mercury, lead and cadmium (Cooley, 2002; John Zink Company, 1991). When chlorinated compounds are converted over catalysts (Dumitriu and Hulea, 1997; Sinquin et al., 2001) chlorine may block the active sites and HCl may react with the carrier and can provoke its crystalline structure collapse, or metal-chlorine species may be vaporized (Spivey and Butt, 1992). A VOC stream containing over 10 ppm of halogenated compounds degrades the activity of a typical catalyst by up to 43% (Süd-Chemie Prototech; John Zink Company, 1991). There are some reviews (Spivey and Butt, 1992; Wilde and Anders, 1994) about the effects of chlorinated compounds on different catalysts, noble metals and metal oxides. Also, there are some reports (Chatterjee, 1992), which evidenced that zeolites are more efficient catalysts for chlorinated hydrocarbons oxidation. A good catalyst for chlorinated compounds oxidation must not be affected by the above mentioned degradation phenomena. In addition, it must also limit the formation of hazardous volatile organics like phosgene, chlorinated dioxins or dibenzofurans and inorganic like metal chlorides or oxychloride (Singuin et al., 2001; Weber et al., 2001). Now, catalyst suppliers offer suchlike catalysts (John Zink Company, 1991). When the organics contain silicone, phosphorus or sulphur their oxides are formed, which cover the active sites decreasing the catalytic activity. Therefore the researchers are looking for catalyst formulations which display a lower affinity for these

metallic oxides.

When the active sites and/or the pores of the catalyst are blocked by a material which is deposited on it is said that the *fooling* of the catalyst occurs (Bartholomew, 1984). The most common foulant is coke (carbonaceous deposits) (Biloen et al., 1977). It is formed by undesired reactions of the organic compounds over the catalyst and its chemical nature varies widely, from graphitic carbon to condensed polymers. Zeolites are the catalysts often compliant with suchlike deactivation.

Catalysts can operate at reasonably high temperatures. In time, *thermal deactivation* can occurs (Trimm, 1991). This is a more problem, in this case the deactivation being irreversible. The high temperature may induce solid state reactions between components of the catalyst; if the active material reacts with the carrier formation of compounds of lower or negligible activities occurs. At high temperature crystals grow or components of the catalyst recrystallize, such processes leading to a loss of surface area. This last form of deactivation is called *sintering* and this involves a redistribution of a material in the solid state in order to decrease the surface energy. The third form of thermal deactivation is represented by catalyst volatilization. It is important to develop low-temperature catalysts, chiefly one of the aims of the researchers is minimization the cold-start period.

Thus, the implementation of catalysts poses some problems, which are challenges for researchers. Besides a great activity (if it is possible even during the cold start-up phase), a good catalyst have to be resistant to the varying temperature conditions and to poisonings, and last but not the least, it must have a low-cost.

The objective of this work is reviewing some studies made on different catalysts on VOCs oxidation, dwelling on innovations with positive consequences concerning the catalytic activity, cost and the durability of the catalyst.

4. Catalytic systems used on VOCs oxidation

4.1. Carriers used on VOCs oxidation

The carrier must be a microporous refractory with a high surface area (between 150 and 300 m^2/g of support) (Cooley, 2002) able to attach the active phase and attract reactants via adsorption. Some of the suchlike materials used on VOCs oxidation are depicted further.

The most commonly used carrier for the active phases for VOCs destruction is γ -Al₂O₃ (alumina) (Müller, 1993). γ -Alumina may have a surface area of 100 to 200 square meters per gram (m²/g) (Dumitriu and Hulea, 1997). It is known that at elevated temperatures its crystal structure changes by sintering, the most thermodynamically stable phase being α -Al₂O₃, with a surface area of 1-5 m²/g. Besides the diminution of the surface area, these "macroscopic movements of the substrate" are accompanied by a decrease in the mechanical

strength of the catalyst (Shkrabina et al. 1995). At the same time, the metal particles dispersed on alumina can stick together (Miyoshi et al., 1989) or/and can be encapsulated into the carrier (process known as the"earthquake phenomena") (Tucci, 1982). Thus, the metal surface area decreases and the catalytic activity is reduced. Moreover, it has been observed that this phase transformation of γ -Al₂O₃ into α -Al₂O₃ is accelerated by the active phase such as Pt (Burtin et al., 1987), and oxides of metals such as Mn, Fe, Ni, Mo, Co, V and Cu (Peivan et al., 1995; Ozawa et al., 1996a), which due to their higher concentrations have a stronger effect than noble metals. Ferrandon made a review (Ferrandon, 2001) of the ions (Zr, Ti, La, Ce, Ba, Sr, Si, Sn, Mg, Ca), which doping into the alumina stabilize it and claimed that by addition of a stabiliser the sintering effects can be inhibited. One of the best additives for inhibiting the Al₂O₃ appears to be La (Burtin et al., 1987; Peiyan et al., 1995), especially when active species are supported on it (Ozawa et al., 1996b; Shkrabina et al., 1995;). A suchlike alumina combined with such oxides is used as washcoat for three-way catalysts (Trovarelli, 1996). Concerning the mechanism of stabilization and the other positive effects of La doping, Ferrandon made a review (Ferrandon, 2001) and has concluded that responsibles are the perovskite LaAlO₃ and the lanthanum- β -alumina.

The other disadvantage of alumina is its low resistance against hydrogen chloride and chlorine (Kulażyński et al., 2002). This is a reason to search for new support materials that are not affected by these products. Above it was already mentioned that now exists suchlike materials. TiO₂-SiO₂ exhibits a high resistance against HCl (Kulażyński et al., 2002) and *Sud-Chemie Prototech's* (Süd-Chemie Prototech) offers specially formulated washcoat and supports which prevent halogen attack and remain stable.

Due its thermal and chemical stability, coupled with acid-base properties and oxygen storage capacity (Cimino et al., 2002), ZrO_2 is meeting with increasing interest in the field of heterogeneous catalysis such as active phase or support/promoter. By supporting noble metals (Fuji et al., 1987; Moles, 2001), metal oxides (Fuji et al, 1987; Ismajilov et al., 1998) and perovskites on zirconia are obtained good catalysts for oxidation. By introduction of zirconia in the ceria lattice (Bozo et al., 2000) a higher mobility of surface and bulk lattice oxygen species is obtained so that methane can be oxidized (its conversion starts at 400°C and is complete at 800°C). Obviously, when platinum was dispersed on this carrier a better activity versus methane oxidation was achieved (conversion starts at 200°C).

Due its properties, SnO_2 can be used like support and like active phase in VOCs catalytic oxidation, also. It has been reported that by doping SnO₂ with SO₄²⁻ anions (Ishikawa et al., 1994) it is obtained a good carrier for Pd on methane oxidation, because of the superacid sites, which impede the sintering, and of the more active oxygen species.

Niobia is another metal oxide used like support in VOCs oxidation, it supporting metals and vanadium oxide (Huuhtanen and Andersson, 1993).

Applications of *zeolites* in environmental catalysis are described in

recently published review articles (e.g., Iwamoto, 1994). The most used for catalytic combustion are metal-containing zeolites (Kucherov et al., 1998). Because their well-defined framework and their acidic properties, which can be controlled, zeolites present some advantages as compared to the other supports, chiefly they have catalytic activity, also. Zeolites can control the type and the particle sizes of the metal supported on them (Kühl, 1999) and can provide high dispersion for the active components (Kalantar Neyestanaki et al., 2000). The intracrystalline space of zeolites has a main role in process kinetics, also (Botavina et al., 2001). Zeolites can be used for a wide range of fuel-air ratios and need lower operating temperatures of VOCs oxidation. Total conversion can be achieved between 140°C and 360°C, depending on the nature of the organic compound (Kullavanijaya et al., 2000; Lahousse et al., 1998). During VOCs oxidation carbonaceous deposits (coke) are formed inside the pores and on the outer surface of zeolites (Derouane, 1985). This leads to their deactivation.

There are very few reports on the use of *pillared interlayered clays* (*PILCs*) on the complete oxidation of VOCs. These are high surface area microporous materials with the diameter and acidity of the porous structure comparable to that of zeolites (Gil et al., 1996). Sometimes, mesoporosity could be also formed. The Brőnsted acidity and the specific surface area of PILCs contributed to a better catalytic performance of the transition metal species, suchlike Al and Cr (Gil et al., 1998) and of the noble metals, like Pt (Gil et al., 2001).

The major drawbacks of traditional metal oxide supports, such as Al_2O_3 , SiO_2 , zeolites, are: the low thermal conductivity - which causes sintering of supported-metal on hot spots of the support during deep and exothermic oxidation of VOCs (Yao et al., 1979), the reaction with water vapour at elevated temperatures and the vaporization of noble metals like oxychlorides (Völter et al., 1987). All these deficiencies cause the attempt of non-oxide materials as supports of catalyst used on catalytic oxidation of VOCs.

Pt supported on *activated carbon* have exhibited several advantages over the traditional Pt/γ -Al₂O₃ (Wu et al., 2000) on VOCs oxidation, such as lower temperature for operation, a more hydrophobic surface, a more chlorine resistance and negligible metal-support interactions. But, activated carbon can not be used at temperature higher than 250°C.

For *hexagonal boron nitride (h-BN)*, a new non-oxide, with a lot of advantages like support (Niedenzu et al., 1965), the temperature is not a problem (Cofer and Econ, 1995). It has a high thermal stability against volatilization up to 800° C in air. Pt supported on h-BN oxidizes more than 90% of gasoline vapour near to 200° C (Wu et al., 2001). In addition, this catalyst reported activities which can be maintained and endure longer than those of Pt-Al₂O₃. "The use of h-BN will become a frontier in catalysis".

4.2. Active phases on total oxidation of VOCs

There are two types of catalysts used in VOCs oxidation (Gallardo-Amores et al., 1998; Papaefthimiou et al., 1997; Zwinkels et al., 1993):

- (i) noble metals and
- (ii) oxides of transition metals.

By reviewing various aspects of the technology (Heck and Farrauto, 1995; Lahousse et al., 1998b) these have been considered the preferred catalysts.

Before discussing performances of each of them some notifications must be made. Firstly, it is important to know that VOCs compounds have different reactivity. The following order of reactivity was established (Hermia and Vigneron, 1993; O'Malley and Hodnett, 1999): alcohols > aldehydes > aromatics > ketones > alkenes > alkanes. Therefore for test the activity of a certain class of catalysts an alkane compound is a good choice (Spinicci et al., 2001) and probably this is the explanation for majority of studies which are made on CH₄. Secondly, in practice the gas stream which must be treated consist in a mixture of organic compounds and that makes the choice of the right catalyst a big problem. It is reported that it is not easy to predict the mixture behaviour from the oxidation of single compound (Barresi and Baldi, 1993; Burgos et al., 2002; Hermia and Vigneron, 1993; Larsson and Andersson, 2000; Musialik-Piotrowska and Syczewska, 1989; Papaefthimiou et al., 1998; Yao and Kummer, 1973; Yao, 1973) because the performance of a catalyst is affected by the mutual effects. Some authors have reported that when VOCs are in mixture inhibiting effects can occur (Botavina et al., 2001; Burgos et al., 2002; Ordóňez et al., 2002; Yao and Kummer, 1973; Yao, 1973). Other studies claim that some VOCs molecules can be activated when are in mixture with other molecules (Hermia and Vigneron, 1993; Musialik-Piotrowska and Syczewska, 1989; Neyestanaki et al., 2000; Völter et al., 1987;) and also selectivity changes could occur (Barresi and Baldi, 1993; Musialik-Piotrowska and Syczewska, 1989). Consequently, it is very important to understand all the factors which determine VOCs reactivity over oxidation catalysts.

4.2.1. Noble metals like active phases for VOCs oxidation

Noble metals are known like catalysts with high activities for oxidative removal of VOCs from gaseous stream. *Pt* and *Pd* are the most commonly used in practice, due to their high intrinsic activity (Papaefthimiou et al., 1997). They are frequently alloyed with other metals such as Ru, Rh, Os and Ir and supported on oxides (Al_2O_3 , SiO_2) (Ordóňez et al., 2002; Vigneron et al., 1996). Because oxidation reactions occur very fast on noble metals these active phases represent only a fraction of percent of the catalyst.

The company *Catalytica* has published an extensive review of the literature prior to 1993 (Horsley, 1993). The activity of metal catalysts depends considerably on the nature of the organic compound. Noble metals are widely applied to gases containing hydrocarbons and oxygenated compounds.

Concerning hydrocarbons, these metals (Pt, Pd, Rh) supported on alumina, silica and zeolites have intensively studied and successfully used (Guisnet et al., 1999; Zwinkels et al., 1993). Excepting for palladium in the case of methane, platinum is recognised to be the most active metal for hydrocarbons oxidation (Burch and Hayes, 1995; Kang et al., 1994). Information concerning the activities of noble metals for oxidation of halogenated hydrocarbons varies in the literature; Petrosius et al. (Petrosius et al., 1993) claimed that noble metals are not choices for this reaction because of their price and their sensitivity of Cl_2 and HCl. In the same year Windawi and Wyatt (Windawi and Wyatt, 1993) claimed that these catalysts have been successfully commercialized for destruction of halogenated compounds, because of their efficiencies, and there are reports which can advocate this affirmation (Lester, 1999; Müller et al., 1993; Petrosius et al., 1993; Spivey and Butt 1992; Windawi and Wyatt, 1993; van den Brink et al., 1998).

A shortlist of some of the studies made on VOCs oxidation with noble metals (Pt and Pd) like active phases is given in Table 1.

Pd is more sensitive to sulphur and lead compounds (Kang et al., 1994; Hegedus et al., 1979) than Pt. Moreover, once sulphur was removed from the gas stream, Pt has a quicker recovery (Beck and Sommers, 1995). It was found that when Pt or Pd is used for hydrocarbons oxidation the influence of sulphur is insignificant especially at high temperature (Beck and Sommers, 1995).

On contrary, on propane oxidation over $Pt/\gamma Al_2O_3$, SO_2 has exhibited a promoting effect (Burch et al., 1998; Ishikawa et al., 1994; Marecot et al., 1994a). The same effect was reported also on naphthalene conversion over Pd/HZSM-5 and over Pd-Pt/HZSM-5 (Kalantar Neyestanaki et al., 2000). The poisoned catalysts can be regenerated at 600-800°C (Ball and Stack, 1991; Beck and Sommers, 1995).

It is important to use a low concentration of noble metals in catalysts as feasible, because of the high cost and limited availability. That means the active metals must be kept at a high degree of dispersion. But, it is reported that at a given amount of platinum or palladium an optimum particle size exists (Marecot et al., 1994b). For example, for propane oxidation this optimum corresponds to a metallic dispersion near 20% on platinum and 50% on palladium, respectively. Some studies (Gandhi and Shelef, 1987; Garetto and Apesteguia, 2000; Hicks et al., 1991; Labalme et al., 1996; Marecot et al., 1994a; Papaefthimiou et al., 1998) claimed that oxidation of hydrocarbons proceeds rather on larger metal crystallites, oxidation being generally considered to be a structure-sensitive reaction, albeit there are some controversies in literature (Chin and Resasco, 1999). In recent works (Pliangos et al., 1997; Papaefthimiou et al., 1998; Garetto and Apesteguia, 2000) it was found that platinum interacts with the carrier affecting the performance of the catalyst. Some authors (Yao et al., 1979) have reported that these interactions bring a negative influence on the catalytic activity.

Table 1. Overview of some studies made on VOCs oxidation with noble metals as catalysts

VOCs			CATALYSTS / REFERENCES			
			Pt - CATALYSTS	Pd - CATALYSTS		
		methane	Bozo et al., 2000; - CeO ₂ -ZrO ₂ [*] . Garetto and Apesteguia, 2000; -	Wang and Xie, 2001b; - $SnO_2 (SO_4^{2-})$.		
			Al_2O_3 .	Maríant et al. 1004h		
	kanes	butana	Marecol et al., 1994.	Marecot et al., 1994b.		
BONS		butane	Labalme et al., 1996.			
	All	avalanantana	Garatta and Anastaguia 2000:			
		cyclopentalle	Al ₂ O ₃ .			
CAR		cyclohexane	Mochida et al., 197 zeolite Y.			
S S	s	ethylene	Pliangos et al., 1997; - TiO_2 , ZrO_2 , γ -Al ₂ O ₃ , SiO ₂ .			
Q	ene	propene	Bozo et al., 2000; - CeO ₂ -ZrO ₂ .			
H	VIK		Marécot et al., 1994b.			
	ł	cyclohexene	Kullavanijaya et al., 2000); - CeO_2 -Al ₂ O ₃ .		
	s	benzene	Hermia and Vigneron, 1993;			
	atic		Sawyer and Abraham, 1994;			
	ü		Papaefthimiou et al., 1998.			
	Arc		Papaefthimiou et a	al., 1997.		
	7	toluene	Burgos et al., 2002;	Paulis et al. 2000; - Al_2O_3 .		
1 - DCs	trichloroethylene		van den Brink et al., 1998.	Spivey and Butt, 1992; Petrosius et al., 1993.		
			Kulażyński et al., 2002; - TiO ₂ -SiO ₂ .			
	s	2-propanol	Burgos et al., 2002;			
	lod		Sawyer and Abraham, 1994;			
	lco	<i>n</i> -butanol	Hermia and Vigneron, 1993.			
	5		Papaefthimiou et al., 1997.			
ŝ	S	acetone	Burgos et al., 2002; Gil et al., 2001; - Al-pillared			
g VOC	one		smeetite.			
	ket	methyl-ethyl-	Gil et al., 2001; - Al-pillared			
nin	-	cetone	smectite.			
tair			Burgos et al., 2002.			
ont		.1 1	Sawyer and Abraham, 1994; -			
Ĵ J		ethyl-acetate	Al ₂ O ₃ . Tichener and Polozzala, 1087:			
0			Papaeffhimiou et al. 1007			
	ers		Papaeffhimiou et al., 1997.	al 1007		
	est	1. 4. 1 4. 4.				
		<i>n</i> -butylacetate		Svezewska, 1989.		
		methyl-	Kullavanijava et al., 2000; - CeO ₂ -Al ₂ O ₃ .			
		methacrylate				
×	diethylamine		Kullavanijaya et al., 2000; - CeO ₂ -Al ₂ O ₃ .			
⁷ 0C		thiophene	Kullavanijaya et al., 2000; - CeO ₂ -Al ₂ O ₃ .			
er V	gasoline vapor		Wu et al., 2001; - h-BN.			
othe		other VOCs	O'Malley and Hodnett, 1999; -	Burck et al., 1995; -		
1		1	moruenne, sinca, aiunnia.	ni203.		

Concerning the thermal resistance, Pd was said to be more resistant to thermal sintering in an oxidizing environment than Pt (Heck and Farrauto, 1995; Hegedus et al., 1979; Spivey and Butt, 1992). Pt volatilizes under oxidizing conditions (Gandhi and Shelef, 1987). Pd can be dispersed as oxides on Al_2O_3 at higher temperature (750-850°C) than does Pt (Kummer, 1986). The interaction between PdO and Al_2O_3 gives considerable activity to Pd- Al_2O_3 catalysts in an oxidising atmosphere. The oxides of Pt and Pd formed during the process are not as volatile in contrast to RuO_2 , OsO_4 or Ir_2O_3 which are also poisonous (Cotton and Wilkinson, 1988).

In spite of their high activities, because of the cost, limited availability and sensitivity to high temperature and poisoning, researchers have been motivated to try the substitution of noble metals and the materials tried as catalysts for VOCs oxidation are: supported transition metals, oxides (simple, complex and combined) of these metals and combination noble metals-metal oxides.

4.2.2. Transition metals like active phases on VOCs oxidation

Several studies have been carried out in order to investigate transition metals supported on different materials on VOCs oxidation. A shortlist of suchlike studies is given in Table 2.

Until recently, because of its chemical inertness (Schwank, 1983) and the difficulty to obtain a high dispersion on common support materials, *gold* was considered as one of the least catalytically used transition metal. In the last decade nanoparticles of gold, dispersed on metal-oxides such as Fe₂O₃, Co₃O₄ and Mn₂O₃, have been managed (Centeno et al., 2002) with a high catalytic activity on VOCs oxidation. When all the metals of the I B group supported on Fe₂O₃ (Scirè et al., 2001) were studied on oxidation of VOCs the following order of the catalytic activity was established: $Au/Fe_2O_3 >> Ag/Fe_2O_3 > Cu/Fe_2O_3 >$ Fe₂O₃. It was found that Au/Co₃O₄ (5%Au) is 70 times more active than (0, 5%) Pt/Al₂O₃ on oxidation of CH₂Cl₂ (Chen et al, 1996), without any noticeable deactivation.

In the last decade transition metals exchanged or impregnated zeolites have been investigated on VOCs oxidation. It has found that transition metal cations increase the zeolite acidity and oxygen chemisorption thus improving the zeolite activity for hydrocarbons conversion (Chatterjee et al., 1992). Metal loaded zeolites (e.g. *Co*-Y, *Cr*-Y, *Cr*-ZSM-5) were also studied on chlorinated compounds oxidation (Kiebling et al., 1998).

4.2.3. Metal oxides like active phases on VOCs oxidation

Metal oxides are an alternative to noble metals as catalysts for VOCs oxidation, although they can have the same catalytic activity but at higher temperature. The most active oxides have a *p*-semiconductor nature and the most frequently used ones are the oxides of *Ag*, *V*, *Cr*, *Mn*, *Fe*, *Co*.

VOCs		OCs	CATALYSTS	REFERENCES	
	alkanes	<i>n</i> -hexane	Au/CeO ₂ /Al ₂ O ₃ and Au/Al ₂ O ₃	Centeno et al., 2002.	
BONS			Group VIII/support	Papaefthimiou et al., 1997; Lahousse et al., 1998b.	
		benzene	Au/CeO ₂ /Al ₂ O ₃ and Au/Al ₂ O ₃	Centeno et al., 2002.	
CAR	s		CuY	Papaefthimiou et al., 1997.	
ŏ	ene		Group VIII/support	Lahousse et al., 1998b.	
DR	alk	toluene	Cr; Co; Cu; Ni/Y zeolite	Antunes, 1999.	
Ξ			Au/Fe ₂ O ₃	Minicò et al., 2001.	
H			Au (Ag, Cu)/Fe ₂ O ₃	Scirè et al., 2001.	
			CuNaHY	Antunes et al., 2001.	
	other hydrocarbons		Transition metals/zeolite	Chatterjee et al., 1992. Neyerstanaki et al., 1995.	
		methanol	Au/Fe ₂ O ₃	Minicò et al., 2001.	
			Au (Ag, Cu)/Fe ₂ O ₃	Scirè et al., 2001.	
		ethanol	Au/Fe ₂ O ₃	Minicò et al., 2001.	
Ő	slo	2-propanol	Au/CeO ₂ /Al ₂ O ₃ and	Centeno et al., 2002.	
N N	oho		Au/Al ₂ O ₃		
Í	alc		Au/Fe ₂ O ₃	Minicò et al., 2001.	
0			Au (Ag, Cu)/Fe ₂ O ₃	Scirè et al., 2001.	
		butanol	Group VIII/support	Papaefthimiou et al., 1997;	
				Lahousse et al., 1998b.	
		methylene	CuCl	Lago et al., 1996.	
	ouo	cloride	Au/Co ₃ O ₄	Chen et al., 1996.	
Cs	drocarb	chlorobenzene	Cr/doped pillared bentonite clays	Storaro et al., 1995.	
l – VC	ed hy		Uranium based catalyst	Hutchings et al., 1996.	
C	chlorinet	chlorinated hydrocarbons	Co-Y, Cr-Y, Cr-ZSM-5	Greene et al., 1996; Ramachandran et al., 1996; Chintawar and Green, 1997.	
$\mathbf{O} - \mathbf{VOCs}$ a		acetone	Au/Fe ₂ O ₃	Minicò et al., 2001.	
		ethylacetate	Group VIII/support	Papaefthimiou et al., 1997; Lahousse et al., 1998b	
other VOCs			Au/support	Bond and Thompson, 1999; Sakurai and Haruta, 1995.	
			Au/Fe ₂ O ₃ Minicò et al., 2000.		
			Au	Centeno et al., 2002.	

 Table 2. Overview of some studies made on VOCs oxidation with transition metals like catalysts

Ni and Cu (Pradier et al., 2000). Some combinations of oxides have exhibited higher activities as compared to single components or even comparable with that of a noble metal. Such catalysts include *Cu-Mn* (Mehandjiev et al., 2000), *Cu-Cr* (Vass and Georgescu, 1996), *Cu-V* (Ahlström and Odenbrand, 1990), *Mn-Ni* (Mehandjiev et al., 1998), *Ag-Co* (Luo et al., 1998), *Cr-Co* (Vass and Georgescu, 1996) *Co-Zn*

(Klissurski and Uzunova, 1993), and *Cu-Al* (Peiyan et al., 1987). The *perovskites*, an interesting structural combination between cations and species of oxygen, are also, interesting like active species on VOCs oxidation. Their performances are discussed later. Transition metal oxides have been found to be very active, both in complete and selective oxidation of hydrocarbons (Anderson, 1986). Several mixed metal oxide systems were applied (Spivey and Butt, 1992; Zwinkels et al., 1993; Petrosius et al., 1993; Wilde and Anders, 1994; Kulazynski et al., 2002) on the oxidation of chlorinated hydrocarbons. Formation of toxic polychlorinated by-products (Froeze and Hutzinger, 1996) and loss of catalytic material as volatile oxychlorides (Griesbaum et al., 1997) are often encountered as drawbacks. An exception seems to be *uranium oxide* (Hutchings et al., 1996). A shortlist of some of the studies made on VOCs oxidation with metal oxides like active phases on VOCs oxidation is given in Table 3. Among the oxides mentioned in the literature, a few seems particularly promising as follows.

VOCs		CATALYSTS		REFERENCES		
			Mn-based	Zhang et al., 1989; 0	Ozkan et al., 1990;	
	kanes	methane	oxides Lahousse et al., 1997; Milellea et al., 1998;			
				Gallardo-Amores et	al., 1999.	
			Sn-based	Meons et al., 1997;	Wang and Xie, 2001a; Wang	
			oxides	and Xie, 2001d; Wa	ng and Xie, 2001c.	
		ethane	Cr ₂ O ₃ /γ-Al ₂ O ₃	; Co ₃ O ₄ / γ-Al ₂ O ₃ .	Kang and Wan, 1994.	
	all	propane	SnO ₂ -La, Pr, N	Nd, Sm, Gd;	Harrison et al., 1999.	
			SnO ₂ -Cr.			
		<i>n</i> -butane	V_2O_5 , Nb_2O_5 , T_1	iO ₂ , MoO ₃ , WO ₃ ,		
			ZrO_2 , HfO_2 , Dy	$_2O_3$, MnO ₂ , Cr ₂ O ₃ ,	Pradier et al., 2000.	
			CuO and Co ₃ O ₄			
NS	Alkenes		Mo-Nb ₂ O ₅		Thorsteinnson et al., 1978.	
BO		benzene	H-ZSM-5		Anderson et al., 1991.	
R	omatics	toluene	V_2O_5/Nb_2O_5		Huuhtanen and Andersson,	
C					1993.	
SO			H-ZSM-5		Dong et al., 1991.	
Ĩ.			Co ₃ O ₄ -Sm		Imamura, 1985.	
H		methyl-	V ₂ O ₅ -Nb ₂ O ₅		Sharipov et al., 1972.	
	aro	substituted				
		aromatic				
		compounds				
		phenanthrene	Mn-Zr mixed	oxides	Fernández López et al.,	
					2001.	
	other hydrocarbons		ZrO_2 - CeO_2		Terribile et al., 1999.	
			SnO-Cr ₂ O ₃		Fabritchnyl et al., 1992;	
					Harrison et al., 1999.	
			SnO ₂ -Cu, Pd,	Cr, Sb	Harrison, 1989.	
			NiO		Yao and Kummer, 1973.	
			α -Cr ₂ O ₃		Yao, 1973.	

 Table 3. Overview of some studies made on VOCs oxidation with transition metal oxides like catalysts

Destruction of volatile organic compounds by catalytic oxidation

	cloromethane		SnO ₂ -Cr ₂ O ₃		Solymosi et al., 1995.			
s	clorobenzene			Y-type zeolitE			Becker et al., 1995.	
				Uranium oxide			Hutchings et al., 1996.	
				Metal oxides			Lago et al., 1996;	
	other Cl-V	VOC				Kiess	ling et	al., 1998;
õ			Ch	Characteristic and common articles			Busca et al., 1998.	
$\tilde{\mathbf{A}}$			Cn	Chronnum and copper oxides			Agarwar et al., 1992, Griesbaum et al. 1997	
J			Cr	Cr_2O_2/Y			Chatteriee and Greene,	
			_	2 5		1991;	1991; Chatterjee et al.,	
						199		
			zec	olites		Chatt	erjee ar	nd Greene,
						1991; Chatterjee et al.,		
	alcohols methanol		Va	$\Omega_c/Nh_2\Omega_c$		1992. Jehng et al. 1007		
	arconois	isopropanol	V_2	$v_2 O_5 / 1 V O_2 O_5$ V ₂ O ₅ /Nb ₂ O ₅			Kera and Kawashima.	
		loopropullor		1203,110203			1988.	
			Mn-Zr mixed oxides			Fernández López et al.,		
്							2001.	
Õ		acetone	Mn ₂ O ₃			Gandia and Korili, 2001.		
20	Ketones		Ag-Mn/Al ₂ O ₃ ; Ag-Co/Al ₂ O ₃ ; Ag- C_2/Al_2O_3 ; Ag-				Luo e	et al., 1998.
inir	methyl ethyl			Mn_2O_2/Al_2O_2 (SiO ₂)			s et al	2000
nta				Mn ₂ O ₂			ia and l	Zooo. Korili 2001
-00		cetone	1011203		Gund	iu unu i	2001.	
0				V ₂ O ₅ , Nb ₂ O ₅ , Ti	O ₂ , MoO ₃ , W	VO_3 , Zr	D ₂ ,	Pradier et
	Esters	ethyl acetate		HfO ₂ ,				al., 2000.
				Dy_2O_3 , MnO_2 , Cr_2O_3 , CuO and) ₄ .	D II I
				$Cr_2O_3/SiO_2(Al_2O_3, MCM-41)$				Pradier et
	Pyridine			$A \sigma - Mn / A l_2 O_2 \cdot A$	σ -Co/Al ₂ O ₂			ai., 2000.
	1 yridine			11g 1011/11/03, 11			1998.	
Cs	other VOCs			MnO ₂ Lahousse et al., 1997		7; Lahousse		
					et al, 1998a.			
0 N				Mn ₂ O ₃	Zhan	ng et al., 1989.		
er				MnO_x/Al_2O_3	Ozka Calla	an et al., 1990.		
Oth			$VIIIO_X/11O_2$	Galla Milei	Gallardo-Amores et al., 1999; Milella et al 1998			
-				Cu(II)Mn(IV)	Yoor	on and Cocke, 1988;		
				O _x				
				SnO ₂	Meor	ons et al., 1997.		

Table 3 (continued)

Chromium oxide containing catalysts (typically 12-25wt. % chromia calculated as Cr_2O_3) seem to have been used broadly and successfully on halogenated compounds destruction (Lester, 1999; Petrosius et al., 1993) and are claimed to be ones of the most active catalysts in destruction of these compounds (Feijen-Jeurissen et al., 1999). The performances of some chromia-based catalysts seem to be superior to noble metals containing catalysts. But, at elevated temperature highly toxic volatile chromium oxychloride ($Cr_2O_2Cl_2$) is formed. This is a reason to restrict the application of Cr-based catalysts to low operation temperatures (Kulazynski et al., 2002).

It was reported that the presence of water can inhibit this unwanted process (Lester, 1999).

Manganese oxides are recognized as being very active for total oxidation of CO and HCs (Zaki et al., 1999; van de Kleut, 1994;) and they are considered to be environment-friendly materials (Reidies, 1986).

According to Zener (Zener 1951), *Mn oxides* suchlike β -MnO₂, γ -MnO₂, Mn₅O₈, α -Mn₂O₃, γ -Mn₂O₃ and α -Mn₃O₄ are optimal surface redox catalysts. In a recent paper (Lahousse et al., 1998a), C. Lahousse et al. have reported that nsutite (γ) form of MnO₂ is more active on VOCs oxidation and in many respects superior to conventional catalysts based on noble metals. By addition of sodium and cesium ions into the structure of Mn₂O₃ (Gandia and Korili, 2001) it was found that its performance on acetone and methyl-ethyl-cetone combustion was improved; in contrast, sulfate ions decrease the activity of Mn₂O₃ and citric acid exhibits no influence.

According to some authors (Moles, 2001), to have Mn-based catalysts very active for oxidation, is needed tetravalent Mn. But, Mn (IV) compounds are usually unstable at high temperature and suffer decomposition to oxygen and less oxidized Mn compounds. Thus, researchers have attempted to stabilize this cation in different compositions analyzed then on VOCs oxidation. Some of them are the following. Mn (IV) stabilized into titania was found to not be very active (Milella et al., 1998). Mn-containing strontium titanate-zirconate perovsites have obtained good results on methane combustion (Daturi et al., 1997); zirconia, well known like good support for metals (Fuji et al., 1987) and oxides (Murakami et al., 1998), stabilizes Mn(IV) cations; several Mn-Zr mixed oxides (Fernández López et al., 2001) oxidize total phenanthrene, as a model for soot (pollutant present in diesel engine waste gases) at quite low temperatures $(250-400^{\circ}C)$ and are very active and more selective to CO_2 on isopropanol oxidation. These materials approach the catalytic activity of pure Mn oxides, but have a higher thermal stability and are more active than Mn-Ti catalysts. Mn-Fe composite oxides have higher activity than precious metal catalyst (Luo et al., 1993) on oxygen-containing organic compounds oxidation. Copper manganate and hopcalite (a mixture of copper (II) and manganese (IV) oxide have exhibited excellent properties like catalysts for post-combustion of model pollutants in wet air (Yoon and Cocke, 1988).

Compared with other metal oxides, for example CuO, Mn oxides react to a lower extent with Al_2O_3 (Strohmeier and Hercules, 1984) to form spinel aluminate, $MnAl_2O_4$, of low activity, and at high temperatures, in presence of steam, present a lower volatility (van de Kleut, 1994). All the above results showed the potential applications of manganese oxides as catalysts for VOCs oxidation. Commercial catalysts based on oxides of Mn are available. They are used in self-cleaning oven walls (Tsyrulnikov et al., 1998).

CuO, a well known component of oxidation catalysts (Park and Ledford, 1998), has exhibited high activity in VOCs emission control applications. CuO/TiO₂ was found to be more active than oxides of Fe, Co and Mn on toluene oxidation (Larsson et al., 1996).

CuO has tolerance to sulphur (Peiyan et al., 1987). Due to its high activity, tolerance to sulphur and refractory nature (Prasad et al., 1984), CuO based catalysts have been considered as substitutes for noble metal catalysts in VOCs control.

It was found that both pure and modified SnO₂ are active oxidation catalysts (Meons et al., 1997). The redox catalytic properties of tin (IV) oxide may be modified substantially by the incorporation of heteroelements (Harrison, 1989) such as copper, palladium, chromium and antimony (for the oxidation of carbon monoxide and hydrocarbons), phosphorus and bismuth (for oxidative coupling and oxidative dehydrogenation) and Fe, Cr and Mn (for methane deep oxidation) (Wang and Xie, 2001a; Wang and Xie, 2001c; Wang and Xie, 2001d). Sn-Cr oxides display markedly improved activity for CH₄ (Solymosi et al., 1995), hydrocarbons (Fabritchnyl et al., 1992) and for CH₃Cl decomposition (Solymosi et al., 1995). The same catalysts proved resistance to SO_2 and water deactivation. A good example of the impact of preparation technique of the performance of a catalyst is the most effective catalyst Cr/SnO₂, produced by impregnation of tin (IV) oxide gel using aqueous chromium (VI) oxide solutions at a Cr:Sn ratio of 0, 15, that, as against a commercial Pt/Al₂O₃ catalyst, exhibits significantly better activity toward propane oxidation (Harrison et al., 1999). Because the performances of chromium and copper-promoted tin (IV) oxides are similar to a Pt/Al₂O₃ these are considered efficacious as three-way emission control catalysts (Harrison et al., 1998). It was reported that incorporation of rare earth cations such as La, Pr, Nd, Sm, Gd into tin (IV) oxide structure showed no significant enhancement of catalytic activity of SnO₂ on propane oxidation (Harrison et al., 1999).

By mixing *silver oxide* with Mn oxide (Ag-Mn(4:1)/Al₂O₃), with cobalt oxide (Ag-Co(4:1)/Al₂O₃) or by addition of small amounts of cerium oxide on it are obtained catalysts with a higher activity than that of individual oxides for CO, acetone and pyridine oxidation (Luo et al., 1998).

 CoO_x is an efficient catalyst for total oxidation reactions (Luo et al., 1998; Ji et al., 2000). In addition it was found that oxides of cobalt can sorb sulphur and CoAlO₄ is stable towards SO₂.

Uranium oxide has been claimed to be an efficient and stable catalyst (Hutchings et al., 1996).

Niobium oxide, besides the role of the support, can be promoter (Luo et al., 1998). It enhanced the activity of molibdate catalyst for alkenes oxidation of vanadium oxide on methyl-substituted aromatic xylene and on o-xylene oxidation (Huuhtanen and Andersson, 1993).

The different roles and applications of *ceria* have been reviewed in several papers (Kašpar et al., 1999; Kummer, 1986; Trovarelli, 1996). Its nonstoichiometric structure (Harrisson et al., 1988), which allows the exchange oxygen with the environment, explains the promoting role of ceria. How ceria acts when it is added at noble metals and at metal oxides catalysts is reviewed by Ferrandon (Ferandon, 2001). Ceria is already used in three-way catalysts (Gandhi and Shelef, 1987; Kummer, 1986), diesel oxidation catalysts (Farrauto and Voss, 1996) and combustion catalysts (Groppi et al., 1999).

An interesting class of oxides is composed of *perovskites.* They are mixed oxides of general formula $ABO_{3\pm\delta}$, where A is usually a lanthanide and/or alkaline earth ion) and B is a transition metal ion (Forni and Rossetti, 2002). Many metals are stable in the ABO₃ perovskite structure (Spinicci et al., 2001). By partially substitution of both A and B, a wide variety of mixed oxides of general formula $A_{1-x}A'_{x}B_{1-y}B'_{y}O_{3\pm\delta}$ could be prepared. The last oxides, owing to their non-stoichiometry, indicated by the δ subscript in the formula, are characterized by structural and electronic defects. The cations A and B are responsible for catalytic activity in different ways (Islam et al., 1996; Leanza et al., 2000; Yamazoe and Teraoka, 1990).

Perovskites, due to their physicochemical and catalytic properties such as ionic conductivity, oxygen mobility within their lattice (Islam et al., 1996), oxygen sorption properties, great stability at high temperature (they are stable in a wide range of temperature, 25-1200°C) and stabilization of unusual cation oxidation states in the structure (Tejuca and Fierro, 1992), were proposed as an attractive alternative to the very active noble metals oxidation catalysts (Forni and Rossetti, 2002).

A short list of some of the studies made on VOCs oxidation with perovskites like active phases is given in Table 4.

So far, perovskites have attracted great interest in hydrocarbons (Fierro, 1993; Tejuca and Fierro, 1992) and other volatile organic compounds (Lintz and Wittstock, 1996) oxidation. Three perovskites - *LaFeO₃*, *LaCoO₃* and *LaMnO₃*-were found to activate the C-H bond cleavage of hexane better than PdO/Al₂O₃ (Spinicci et al., 2001) and to be active on methane combustion (Ciambelli et al., 2001; Tejuca and Fierro, 1992). A catalytic activity comparable to that of Pt/Al₂O₃ catalysts was reported for some perovskites studied on oxidation of methane (Zwinkels et al., 1993) and of chlorinated compounds (Sinquin et al., 1997). Concerning chlorinated compounds, they could be total oxidized by perovskites and in addition, the amounts of formed byproducts are significantly lower (Weber et al., 2001), but this process is connected with deactivation (Kiebling et al., 1998).

Because the preparation method of perovskites involves a high temperature heating step, perovskites have a low specific surface area (<20 m²/g). To increase it and also, to increase the mechanical strength and the lifetime, this active phases are dispersed over refractory supports (e.g. γ -Al₂O₃, MgO) (Zwinkels et al., 1999; Cimino et al., 2001). Dispersion on zirconia provides a significantly enhanced specific activity of perovskites (Fuji et al., 1987; Cimino et al., 2002), while on all of the other oxide supports a remarkable decrease of specific activity per unit mass of perovskite, particularly in the case of Al₂O₃ and SiO₂ (Fuji et al., 1987). In addition, zirconia can be stabilized against sintering at high temperatures, by a variety of dopants (La, Si) (Franklin et al., 1991).

VOCs		CATALYSTS / REFERENCES						
		LaCoO ₃ La(Ce)CoO _{3+δ} [*] ; La(Sr)CoO ₃ / supports ^{**}	$\frac{LaMnO_{3+\delta}}{LaMnO_3/ZrO_2}^*$	La(Sr)B(M n, Fe, Co)O ₃	LaBO ₃ B: Ni, Co, Fe, Mn, Cr; LaCo(Fe)O ₃			
	methane	Forni and Rossetti, 2002.*	Cimino et al., 2002.*	Seiyama,				
	propane	Fuji et al., 1987.		1993;	~			
arbons	<i>n</i> -hexane				Spinicci et al., 2001. Szabo et al., 2002.*			
Hydroc	saturated and unsaturated C ₂ hydrocarbons	Lago et al., 1996; Taylor et al., 2000; Poplawski et al., 2000; Sinquin et al., 2000a; Sinquin et al., 2001.	References of LaCoO ₃ .	References of LaCoO ₃ .				
hlorinated compounds	chlorinated methanes	References of LaCoO ₃ ; Sinquin et al., 1998; Kiebling et al., 1998; Sinquin et al., 2000b.	References of LaCoO ₃ ; Sinquin et al., 1998; Kiebling et al.,1998; Stephan et al., 1999; [*] Sinquin et al., 2000b.	References of LaCoO ₃ ; Kiebling et al., 1998.	Sinquin et al., 1998.			
	chlorinated C ₂ hydrocarbons	Kiebling et al., 1998.	Stephan et al., 1999 [*] ; Sinquin et al.,2001.	Lago et al., 1996; Kiebling et al., 1998.	Kiessling et al., 1998; Stephan et al., 1999.			
0	chlorobenzene	References of LaCoO ₃ .	References of LaCoO ₃ .	References of LaCoO ₃ .				
	other chlorinated compounds	Kiebl	ing et al., 1998; Web	er et al., 2001.				
0- VOCs	oxygenated compounds			Yamazoe and Teraoka, 1990; Seiyama, 1993.				

 Table 4. Overview of some studies made on VOCs oxidation with perovskites like catalysts

Recently, this major problem of perovskites (low specific surface area) was resolved by Kaliaguine et al. (Kaliaguine et al., 2001). The new technique of preparation involved is called "mechano-synthesis" and is performed at essentially room temperature. The samples obtained have surfaces area in excess of 100 m²/g and also, surface chemical properties of a particular interest. Another solution of the problem can be the synthesis of highly dispersed

nanometre-size powders, obtainable by the" flame-hydrolysis technique" (Forni and Rossetti, 2002).

4.2.4. Metal oxides versus noble metals. Their combinations like catalyst for VOCs oxidation

Generally, metal oxides are more sensitive to sulphur than noble metals (Ball and Stack, 1991), but some metal oxides have exhibited a good sulphur resistance. In fact, the higher overall loading of metal oxides in the catalysts, that means a number of active sites orders of magnitude larger than that of noble metal catalysts, makes them more tolerant to poisons than noble metals (Hegedus et al., 1979). This above thing, combined with the higher thermal stability and the price incomparably good of the metal oxides (Spivey and Butt, 1992), and the higher specific activity of noble metals, are considered like different good properties by themselves. A good example that noble metals and metal oxides action different and complementary is the study (Kulazynski, 2002) made on trichloroethylene oxidation with Cr₂O₃, V₂O₅, Pt and Pd; the oxides, supported on TiO₂-SiO₂, have displayed a higher activity (expressed per weight) than noble metals, but noble metals are more stable and can be used under more heavy conditions (higher temperature and higher concentration of chlorine). All these things have persuaded researchers to mix these materials and analyze the mixtures on VOCs oxidation.

It has reported that noble metals and transition metal oxides enhance the activity each other mutually (synergism effect). Ferrandon have made a short review of these interactions (Ferrandon, 2001). Some examples of synergism are: doping Pd with oxides of Co or Cr it is restrained the poisoning effect of sulphur (Khairulin et al., 1997); by incorporation of Pt and Pd into metal oxides the selectivity toward CO_2 is increased (Ménézo et al., 1993); the activity of Pd on CH_4 oxidation is increased by addition of metal oxides, particularly oxides of Ni (Widjaja et al., 1999).

5. Conclusions

The general conclusion from this paper is that the catalysis, besides all its applications, could be a very good tool in the fight against the environment pollution with VOCs, as well.

By the lower temperatures involved for destruction of VOCs, a lot of advantages as against corresponding thermal options are provided, such as: attractive saved energy, longer equipment life and higher selectivity to CO_2 and H_2O . In addition, in our days the limitations involved by the first catalysts used in this process are overcome or minimized. That is due to the aggressive research programme, developed by the catalyst suppliers. The new formulations of the catalysts allow for immunity to deactivators, for a higher thermal stability and a lower cost, as well. Thus, new doors, that could not have been considered a few years ago, have been opened.

In addition, operating a pilot system off of process slip streams and

watching for trends, the optimization of the catalyst selection for the specific waste stream is allowed.

The catalytic oxidation, as an after treatment technology, provides an attractive combination of features.

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